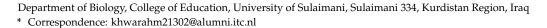




Article

# Mapping Impacts of Climate Change on the Distributions of Two Endemic Tree Species under Socioeconomic Pathway Scenarios (SSP)

Barham A. HamadAmin and Nabaz R. Khwarahm \*



Abstract: Pistacia eurycarpa Yalt and Pistacia khinjuk Stocks are two important endemic tree species inhabiting mountainous regions in Iraq. Their cultural, medical, and ecological benefits have captured the interest of this study. Numerous researchers have revealed how and to what extent global climate change alters species' habitats and distribution. This approach aims to quantify the current and future distribution of these tree species in the region and to provide baseline data on how Pistacia respond to the changing environment. Three socioeconomic pathway scenarios (SSP 126, 245, and 585) in two general circulating models (GCMs), MIROC-ES2L and BCC-CSM2-MR, have been utilized to examine the probable future geographical shift of these species during different time periods (2041–2060, 2061-2080, and 2081-2100). This study used the MaxEnt model and geospatial techniques for: (i) anticipating the present level of distributions and assessing the impact of climate change on these species' possible future distributions; (ii) estimating the areas of species overlap; and (iii) finding the most significant environmental variables shaping their distributions, according to 11 environmental variables and 161 known localities. The findings revealed that 30 out of 36 modeling results showed range expansion in both the MIROC-ES2L and BCC-CSM2-MR models with 16/18 for P. eurycarpa and 14/18 for P. khinjuk. The overall species range expansions and increase in habitat suitability (mainly in the north and northeast) were related to precipitation during the wettest months, topography, and soil type structure (i.e., Chromic Vertisols, Lithosols, and Calcic Xerosols). These recent discoveries provide priceless new information for forestry management efforts and the conservation plan in Iraq, particularly in the overlapping areas in the mountainous regions and highlands. Geospatial approaches and correlation-based modeling are effective tools for predicting the spatial pattern of tree species in the mountain environment.

Keywords: climate change; GIS; Iraq; maxent model; Pistacia; prediction



Citation: HamadAmin, B.A.; Khwarahm, N.R. Mapping Impacts of Climate Change on the Distributions of Two Endemic Tree Species under Socioeconomic Pathway Scenarios (SSP). Sustainability 2023, 15, 5469. https://doi.org/10.3390/ su15065469

Academic Editor: Francesco Sottile

Received: 10 February 2023 Revised: 1 March 2023 Accepted: 17 March 2023 Published: 20 March 2023

Correction Statement: This article has been republished with a minor change. The change does not affect the scientific content of the article and further details are available within the backmatter of the website version of this article.



Copyright: © 2023 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https://creativecommons.org/licenses/by/4.0/).

## 1. Introduction

The genus *Pistacia* (*Pistacia* L. 1753) belongs to the order Sapindales and the Anacardiaceae family. The taxonomy of the genus' eleven species is based on factors such as molecular analysis, morphology of the leaf, and floral structure [1]. *Pistacia* can be observed on sandstone and limestone hills with an elevation range of 30 to 2500 m, steep dry slopes, and rocky hillsides. The genus originated in Central Asia, and according to Mohannad G and Duncan M [2], it is distributed across southern Europe (the Mediterranean area), West and northern Africa, Central Asia, Central America, and the Middle East [3]. Its distribution in the Middle East is especially concentrated in the Kurdistan Region of Iraq (KRI), where it is located between the northeast of Iraq, southern Turkey, northeast Syria, and western Iran [4].

Only three *Pistacia* species, *Pistacia eurycarpa* Yalt, *Pistacia vera* L., and *Pistacia khinjuk* Stocks, have been described and well documented so far, in the KRI [2,3]. Only *P. vera* L., is grown in cultivation; while the other two species are endemic to the KRI's Sulaimani, Erbil, Duhok, and Halabja Governorates (Figure 1). *P. khinjuk* and *P. eurycarpa* (Figure 2) reach

heights of 3 to 7 m and 4 to 20 m, respectively. Additionally, *P. eurycarpa* according to the conclusion of Rankou, et al. [5] is the same as *P. atlantica* subsp. *kurdica*, which is a native Kurdish wild tree plant, which the locals call "Dar qezwan" or "Dareben" [4].

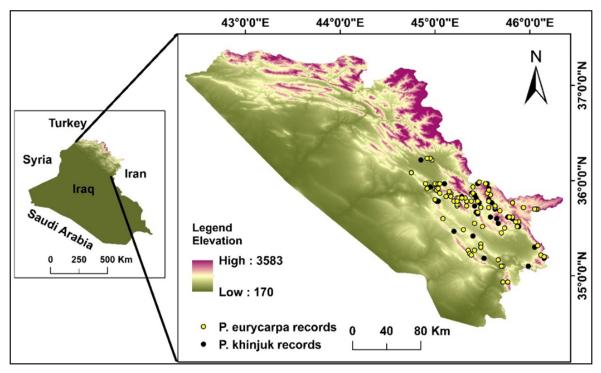


Figure 1. Study area: Kurdistan Region of Iraq and species records.



**Figure 2.** *P. khinjuk* (right) (Aug 2021, plant height 2.5 m) Sharbazher–Kanarwe area. *P. eurycarpa* photo (left) (July 2021, plant height 6 m) Sharbazher–Khamza area, Sulaimani Governorate. Photos by Barham A. HamadAmin.

Currently, in Iraq, habitat degradation [6–8], overexploitation (ruthless collecting by people for domestic needs, such as firewood, medicine, and national trafficking), forest fires [9] and warfare [10] are some of the threats that significantly influence forest trees, including *P. khinjuk* and *P. eurycarpa*. These threats could increase as the climate changes. Global climate change and other factors, including elevation and soil properties, affect numerous types of plant species, particularly their spatial distributions [11]. Not only does climate change affect present plant phenology, physiology, or distribution, but it also affects how the species will fare in the future [11–13]. The target species of this study have remarkable cultural, ecological, and ethnobotanical values. For example, isolated compounds

Sustainability **2023**, 15, 5469 3 of 25

and crude extracts of such species, including flavonoids, oil, resin, tocopherols, fatty acids, phytosterols, and phenolic compounds, have remarkable pharmacological features and medicinal significance (e.g., analgesic, anti-inflammatory, antifungal, antimicrobial) [14], in addition to their important contribution to the stabilization and integration of the ecosystem of the forest and their nutritional qualities [15]. Moreover, locals consume the fruits of *P. eurycarpa*, which has considerable antibacterial action against *Helicobacter pylori*, as well as utilize the resin from the plant to manufacture chewing gum, jam, and food [16], and to aid digestion and stomach relief [17]. Animals in veterinary departments have been treated with the fruits and leaves from *P. khinjuk*, and the resin is used to alleviate nausea, motion sickness, and stomach pain [15].

Previous research on *Pistacia* has mostly concentrated on its medicinal features [16,18,19], ethnobotany [4,14,20], cytology [21,22], classification [2], molecular biology [23–25], morphology [24,26], anatomy [21,27], and to a limited extent, distributions [1,28]. Nevertheless, to the best of the researchers' knowledge, no comprehensive studies about the geographic ranges of the *Pistacia* genus have ever been conducted in Iraq, and none have considered the present and future potential distributions of the *P. eurycarpa* and *P. khinjuk* species. Thus, gathering baseline data on these two tree species in Iraq's mountain ecosystems will help us better understand how climate change is affecting the area and what management measures are working.

In terms of the current methodologies, the most extensively used methods for mapping and assessing species distributions in respect to environmental indicators are geographic information systems (GIS) and species distribution models (SDMs) [29]. Statistical methods are used in SDMs to assess and predict the species spatiotemporal occurrence [30,31]. SDMs estimate the species geographical distribution, namely, the probable locations where the species' presence meets its environmental characteristics [32]. They can also be used to identify locations that might be used to organize future survey efforts, mark protected areas for conservation initiatives, reintroduce species, and provide important assistance in discovering previously undiscovered populations [33]. As a machine learning technique, maximum entropy (MaxEnt) is the least sensitive to small sample sizes that relies on presence-only data, and is one of the most extensively used and well-applied models among the SDMs [34].

In this study we aim to: (i) predict and map the present and future spatial distributions for these species in Iraq; (ii) estimate the distributional variations between the present and future statuses (loss and gain, shift of range); (iii) map areas where the two species' ranges overlap; (iv) identify pertinent environmental factors affecting the species' distribution; and (v) estimate the direction and magnitude of the region shift from the present to the future.

#### 2. Materials and Methods

#### 2.1. Study Area

Iraq is situated in the Middle East, covering an area of 438,320 km², and is bordered by Iran to the east, and Jordan to the west, Turkey to the north, and the Arabian Gulf and the Kingdom of Saudi Arabia to the south (longitude: 38°45 to 48°45′ E and latitude: 29°15′ N to 38°15′ N) (Figure 1). Iraq is divided into four major physiographic regions: the northeastern highlands, the uplands (areas of undulating and hilly terrain in the north between the Euphrates and Tigris), the central and southeast plains of alluvial (marshlands), and the western and southern desert [35,36]. Additionally, it has four distinct seasons: fall (October to November), a hot and dry summer (July to October), spring (March to May), and a chilly and rainy winter (December to February). The driest seasons are summer and autumn, while winter and spring get around 90% of the yearly precipitation [37]. Most of Iraq's Kurdish population lives in the northeast of the country, where the Zagros Mountains rise with several peaks. The mountains and highlands are often inaccessible with an elevation range from 800 to 3544 m [38]. In the desert of the southern region to the mountainous areas of the northeastern region, the average daily mean temperature during

Sustainability **2023**, 15, 5469 4 of 25

the past year and the precipitation ranges from  $28 \,^{\circ}\text{C}$  to  $9.8 \,^{\circ}\text{C}$  and below  $100 \, \text{mm/year}$  to more than  $900 \, \text{mm/year}$ , respectively. The country is prone to drought and heat wave calamities because of the exceptionally high temperatures and near nonexistent rainfall from May to October [37]. Erbil, Duhok, Sulaimani, and Halabja are the four governorates that make up the KRI, which is placed at  $37^{\circ}38' \, \text{N} \, 46^{\circ}35' \, \text{E}$  and has an overall area of  $51,173.65 \, \text{km}^2$ . The KRI is dry and hot in the summer and wet and cold in the winter [39,40]. The KRI is used as the model boundary in this research due to the region's similarity in terms of climate, physiography, and forest structure and composition [39].

#### 2.2. Ground Data for P. eurycarpa Yalt and P. khinjuk Stock

Multiple surveys were carried out in the Sulaimani Governorate between 14 July and 25 December 2021 to obtain the presence data for both species. These data were gathered using the stratum sampling technique [41]. The villagers' expertise and *Flora of Iraq* [3], together with a handbook guide, have all been crucial in identifying and gathering the data. Each sample record was visited, looked through, photographed, and herbarium-ready (Figure A1). Plant phenotypes (observable characteristics) were used in this research to recognize and classify both species. Moreover, 161 GPS ground records (*P. eurycarpa* = 89, *P. khinjuk* = 72) were initially collected as a result of the survey efforts. Based on Boakes, et al. [42], geographical filtering of 1 km was performed to the dataset to minimize the spatial autocorrelation between the sites. This resulted in a reduction of the points of sample to 135 records (n = 78 and n = 57, for both species). Applying the spatial filtering approach is beneficial in reducing sample error and improving the variability of the altitudinal differences among the sites [43]. For the spatial filtering and quality verification, the extended SDMtoolbox 2.5 and ArcGIS 10.3 (ESRI, Redland, CA, USA) software were used [44].

#### 2.3. Environmental Datasets

According to Guest and Townsend [3], both species are frequently found in Iraq's forest zones and sporadically in the forest-steppe transition and moist steppe zones. This is especially true in the northeastern regions, which primarily include the governorates of Duhok, Erbil, Sulaimani, and Halabja. Pistacia does not naturally exist in other regions of Iraq. The KRI's geographical scope was thus, used for our modeling needs to extract the environmental variables, while clipping off the research region (Figure 1). Numerous research has emphasized the significance of a specified geographic range in SDM modeling [45,46]. Climate, topography, and edaphic environmental factors are some of the major factors affecting the species' range [47,48]. Multiple environmental factors have been used to develop the model. To construct the model, 19 bioclimatic variables were taken into account for the present (from 1970 to 2000) and future climatic time frames (i.e., 2041–2060, 2061-2080, and 2081-2100), under three common socioeconomic pathways (SSP 585, 245 and 126) [49]. The present and future climate data were obtained from the world climate database (www.worldclim.org) [50]. The Earth System version 2 for long-term simulations (MIROC-ES2L) and the Beijing Climate Center Climate System Model (BCC-CSM2-MR), two models from the most recent sixth level of the Coupled Model Intercomparison Project (CMIP6), are included in the future climate models [51].

In comparison to earlier models (i.e., CMIP5), the CMIP6 models are very reflective of the concentrations of greenhouse gases (GHG) and provide superior simulations of temperature under atmosphere—biosphere transitions, complex topography, and an improvement in model resolution. In order to comprehend: (i) the future climate in the context of internal variability, predictability, and uncertainty, (ii) the Earth's reaction to forcing, (iii), the origin of the models' systematic biases, and (iv) the cause of the models' systematic biases, the CMIP6 models include specific climatic issues. In CMIP6, the current analysis made use of the recently created future scenarios [52], known as shared socioeconomic pathways (SSPs) that upgrade the depiction of the expected socioeconomic and technological growth in the scenarios of future climatic conditions [53]. Only three of the five SSP scenarios, fossil

Sustainability **2023**, 15, 5469 5 of 25

fuel-based development (SSP5), middle-of-the-road development (SSP2), and sustainable development (SSP1), were actually employed [53].

The topographic variables were collected from the Shuttle Radar Topography Mission (SRTM) (http://srtm.csi.cgiar.org/srtmdata/ (accessed on 1 February 2022)), including the DEM (digital elevation model), aspect (estimated in degree from the DEM), and slope. The remaining edaphic factors, including the soil pH and soil moisture level, were also retrieved from the Center for Sustainability and the Global Environment (SAGE) (http://www.sage.wisc.edu/atlas/index.php (accessed on 10 February)) database. Additionally, the datasets for geology and soil type were obtained from the Food and Agriculture Organization of the United Nations (https://www.fao.org/soils-portal/data-hub/soil-maps-and-databases/ (accessed on 1 March 2022)) and the United States Geological Survey (https://certmapper.cr.usgs.gov/data/apps/world-maps/ (accessed on 1 March 2022)).

The ArcGIS 10.3 software was utilized to pre-process the data and spatially resample all the variables to a resolution of 1 km [54]. Due to the great spatial correlation (collinearity) among the variables, for both species, out of the 27 environmental parameters only 11 were involved in the model construction (Table 1). To keep predictors with a pairwise Pearson's correlation of |r| 0.8, a conditional threshold technique was used in order to avoid collinearity [55]. For the predictors, the pairwise Pearson's correlation analysis was performed using the SDMtoolbox extension and ArcGIS 10.3. (ESRI, Redland, CA, USA) [56].

**Table 1.** The environmental factors taken into account for modeling. (The model development process employed just the variables in bold).

Variable	Code and Unit	P. eurycarpa	P. khinjuk
Annual mean temperature	Bio1 (°C)		
Mean diurnal range	Bio2 (°C)	$\sqrt{}$	$\checkmark$
Temperature seasonality	Bio4 (standard deviation $\times$ 100)		
Isothermality (BIO2/BIO7)	Bio3 (×100)		
Max temperature of warmest month	Bio5 (°C)		$\checkmark$
Min temperature of coldest month	Bio6 (°C)		•
Temperature annual range	Bio7 (Bio5-Bio6) (°C)		
Mean temperature of wettest quarter	Bio8 (°C)		
Mean temperature of driest quarter	Bio9 (°C)		
Mean temperature of warmest quarter	Bio10 (°C)		
Mean temperature of coldest quarter	Bio11 (°C)		
Annual precipitation	Bio12 mm	$\checkmark$	$\checkmark$
Precipitation of wettest month	Bio13 mm	$\sqrt{}$	$\checkmark$
Precipitation of coldest quarter	Bio19 mm		
Precipitation of warmest quarter	Bio18 mm		
Precipitation of driest quarter	Bio17 mm		
Precipitation of wettest quarter	Bio16 mm		
Precipitation seasonality	Bio15 mm		
Precipitation of driest month	Bio14 mm		
Slope	Slope (degree)	$\checkmark$	$\checkmark$
Aspect	Aspect (degree)		
Soil type	FAO soil classification	$\checkmark$	$\checkmark$
Soil carbon	Soil carbon (%)	$\sqrt{}$	$\sqrt{}$
Soil moisture	Soil moisture (mm)	$\stackrel{\cdot}{\checkmark}$	$\sqrt{}$
Soil pH	Soil (parts hydrogen)	$\sqrt{}$	$\sqrt{}$
DEM	Digital elevation model (m)	·	·
Geo		$\sqrt{}$	$\sqrt{}$

#### 2.4. Model Building

The MaxEnt model was chosen due to its ability to produce results that are comparable to those of presence-absence model approaches and presence-only dependent model types [34], as well as its superior predictive accuracy when compared with other

Sustainability **2023**, 15, 5469 6 of 25

approaches [32,57]. For P. eurycarpa and P. khinjuk, 70% of the presence data points were used to train the models, while the remaining 30% were used for model validation. In this investigation, the algorithm iterations maximum number was 500, with the choice of model replicate set to 10 (maximum entropy). The 10 model replicate option yields average habitat suitability maps for P. khinjuk and P. eurycarpa. To account for the species' presence records (n = 78 and n = 57, respectively), the background points were adjusted to 570 [30]. The multiplier of regularization was adjusted to 1 [34]. Tuning the regularization multiplier may change how complicated or simple the model is. For example, raising the number above one makes the model simpler, while lowering the value below one makes the model more complex [58]. The jackknife test was also used, for assessing the factors' relative significance and contributions to the likelihood of distribution of species' habitats [59]. Additionally, from 10 model outputs (10 repetitions) the value of average (i.e., the value of the threshold) was utilized to define the likelihood of unsuitable distribution and habitat suitability for these species (occurrences probability) [60]. For P. eurycarpa and P. khinjuk distribution, pixels with values equal to or greater than 0.3 were deemed appropriate regions, whilst pixels with values lower than 0.3 were considered unsuitable areas. According to these threshold values, the suitability maps for the two species were classified as follows: high suitability (0.67–0.89), medium suitability (0.45–0.67), low suitability (0.3–0.45), and unsuitable (0–0.3) [61] (Figure 3). The ArcGIS 10.3 platform's spatial analysis toolbox was employed to conduct these operations.

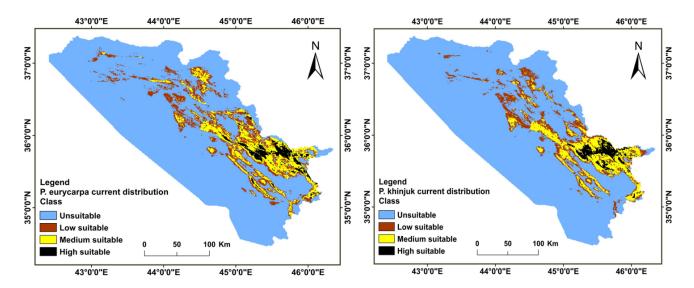


Figure 3. Potential current distribution of both studied Pistacia sp. in the KRI.

#### 2.5. Model Evaluation

One of the most common evaluation metrics for the presence–dependent modeling approach is the area under the curve (AUC). The AUC is one of the most extensively utilized tools to evaluate a model's discriminatory power that assesses how well the model outputs discriminate between locations where observations are present and absent [8]. The AUC values range from 0 to 1. The performance of the model has been divided into the following ranges: outstanding (0.9–1), very good (0.8–0.9), decent (0.7–0.8), acceptable (0.6–0.7), and bad (0.5–0.6) [62]. Other metrics, such as TSS [63], are often used in similar studies, however, TSS and AUC are highly correlated.

#### 2.6. Analysis of the Distribution Change between the Habitat of the Present and Future for the Species

To estimate the transition between the present and future habitat distributions, several spatial techniques in the ArcGIS platform were deployed. Four categories were utilized to categorize the changes in distribution for the species in the KRI: (1) no change (i.e., it is inhabited by the species at the moment, and it is still occupied throughout time frames

Sustainability **2023**, 15, 5469 7 of 25

2041–2060, 2061–2080, and 2081–2100 years); (2) no occupancy (i.e., the locations that are unsuitable now and that will remain unsuitable in the future); (3) expansion of the range (i.e., future habitats that will be suited to the species); and (4) contraction of the range (i.e., future regions that will be lost for every species). For the species over its range (i.e., geographical extent), the centroid was estimated to provide a clear understanding of the changes in distribution between the scenarios of the present and the future. Therefore, the study was carried out using SDMtoolbox, a GIS tool created by Brown, Bennett and French [56], to identify the centroid of the distribution changes in the suitable areas with attributes on the magnitude and direction, by condensing the geographical extent of the presence recordings to a single location (spatial mean).

#### 3. Results

#### 3.1. Performance of the Model

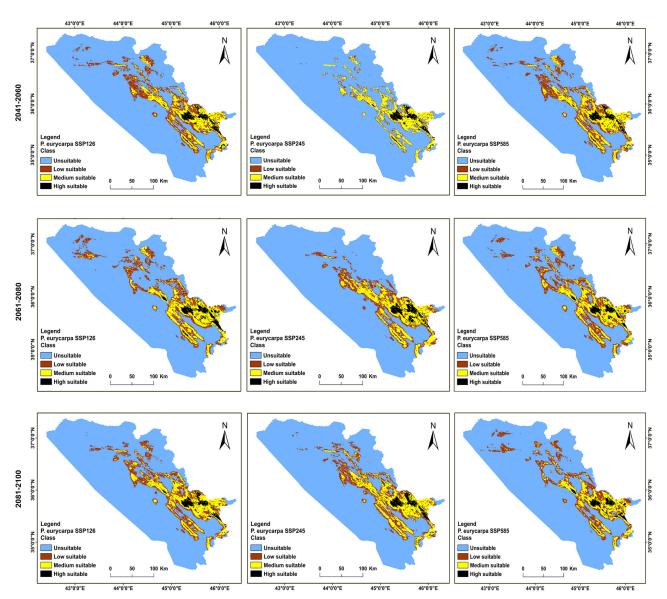
For *P. khinjuk* and *P. eurycarpa*, the AUC values of the model's performance were 0.853 and 0.846, respectively, and the first species showed a stronger discriminatory power in the outputs of the model. Under the chosen environmental circumstances, for the two species, the likelihood of the distributions of habitat suitability as a whole showed a logical effectiveness.

#### 3.2. Distributions of the Habitat in the Present and Future for P. khinjuk and P. eurycarpa

The simulations derived from the MIROC-ES2L and BCC-CSM2-MR climate models, for instance, reveal that the range of *P. eurycarpa* will change from its existing distribution under the SSP 585, 245 and 126 scenarios, during the periods 2041–2060, 2061–2080, and 2081–2100. Particularly, at the cost of the unsuitable sites, the species will gain some territory (grow in certain places). In the BCC-CSM2-MR models, the unsuitable regions grew the most by 3328 km<sup>2</sup> (6.5%) under the SSP 245/2041–2060. While the SSP 585/2041–2060 had the largest area drop, by 605 km<sup>2</sup> (1.18%). Additionally, it is anticipated that the previously acceptable regions will fluctuate in size depending on where they are located (habitat shift). For instance, it is predicted that the category of high suitability will decline in every utilized scenario, but that the overall classes of habitat suitability (i.e., high suitability, medium, and the sum of the low suitability areas) will enhance in all but the SSP245 scenario, which is predicted to decline in the 2041–2060 time period, from the recent  $10,443 \text{ km}^2$  (20.43%) to 7115 km<sup>2</sup> (13.92%). However, it is anticipated that in all the applicable situations, employing the MIROC-ES2L models, both categories of high suitability and habitat unsuitability will diminish. However, for each scenario, the overall classes of suitability (i.e., the high suitability, medium, and the total of the low suitability areas) will rise (Tables A2 and A3; Figures 4 and 5).

Both of the models of the climate show a dynamic shift in the ranges for *P. khinjuk* as well. The two models reveal shifts in the trend of habitat unsuitability, and most classes of overall suitability are anticipated to rise under future climate scenarios. For instance, SSP 245/2061–2080 shows the biggest rise in habitat unsuitability under the BCC-CSM2-MR models by 745 km<sup>2</sup> (1.45%), whereas SSP 245/2041–2060 shows the greatest reduction in habitat unsuitability by 505 km<sup>2</sup> (0.99%). With the exception of SSP 585/2081–2100, and SSP126, 245/2061-2080, all the applied scenarios predict an increase in the total suitability classes, whereas the former scenarios predicted a decrease from 9230 km<sup>2</sup> (18.06%) in the present to 8799 km<sup>2</sup> (0.84%), and 8986 km<sup>2</sup> (0.47%) to 8484 km<sup>2</sup> (1.45%), in the future. The modeling of the MIROC-ES2L results also revealed a progressive shift in P. khinjuk's distribution. For instance, all of the applicable scenarios expect a reduction in habitat unsuitability, with the exception of SSP 245/2041-2060, which predicts an increase of 192 km<sup>2</sup> (0.37%). However, all scenarios predict a rise in the overall classes of suitability (i.e., the total of the high, medium, and low suitability regions), with the exception of SSP 245/2041-2060, which predicts a decline from the present 9230 km<sup>2</sup> (18.06%) to the future 9037 km<sup>2</sup> (17.69%) (Tables A7 and A8; Figures 6 and 7). In conclusion, the BCC-CSM2-MR model shows that the mean of all the eligible habitats will grow for *P. khinjuk*, but contract

for *P. eurycarpa*. The MIROC-ES2L model, however, shows that the mean of the total amount of habitat that is appropriate for both species will increase.



**Figure 4.** Distributions for the future habitat of *P. eurycarpa* under (SSP 585, 245, and 126) climate change scenarios in the BCC-CSM2-MR model in the years 2041–2060, 2061–2080, and 2081–2100 in the KRI.

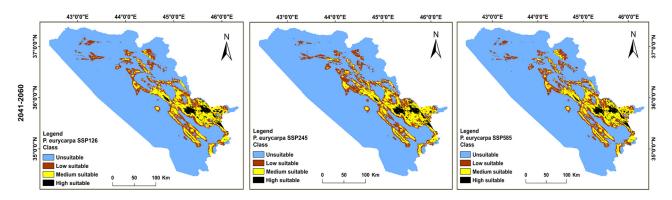
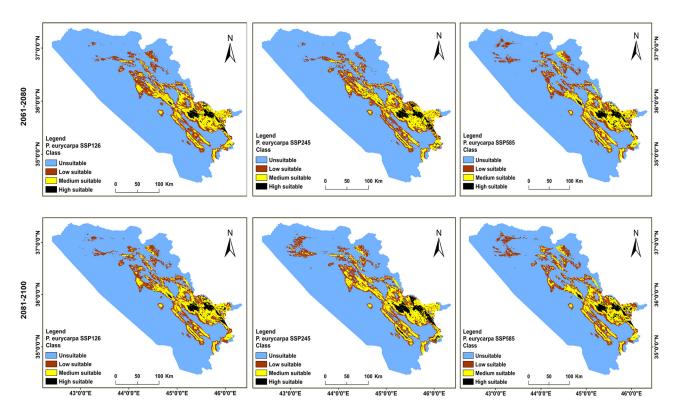


Figure 5. Cont.



**Figure 5.** Distributions of the future habitat of *P. eurycarpa* under (SSP 585, 245, and 126) climate change scenarios in the MIROC-ES2L model for the years 2041–2060, 2061–2080, and 2081–2100 in the KRI.

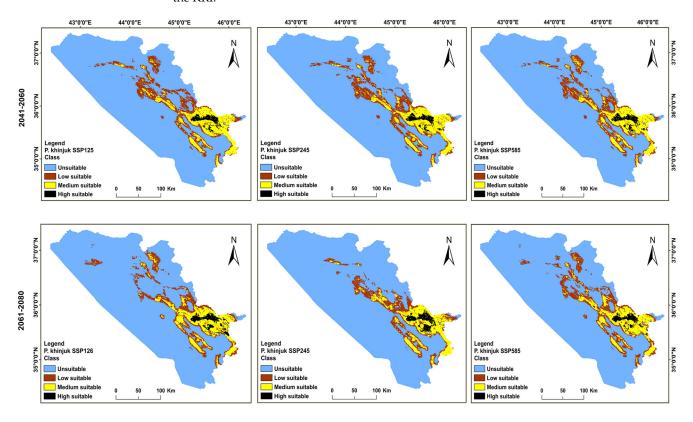
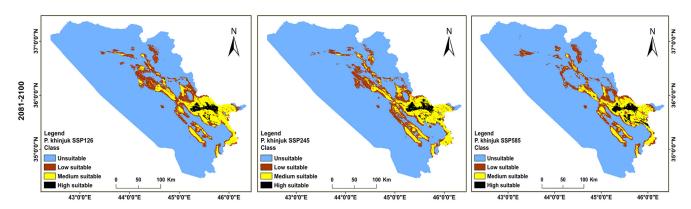
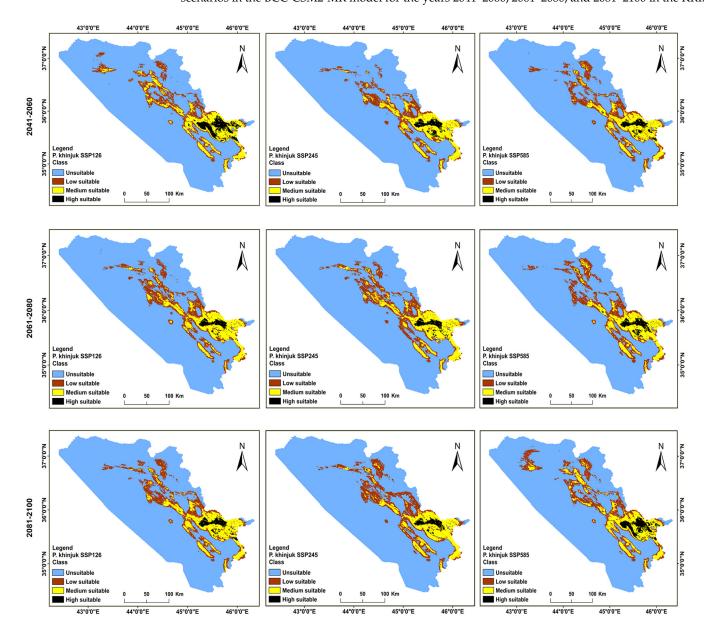


Figure 6. Cont.



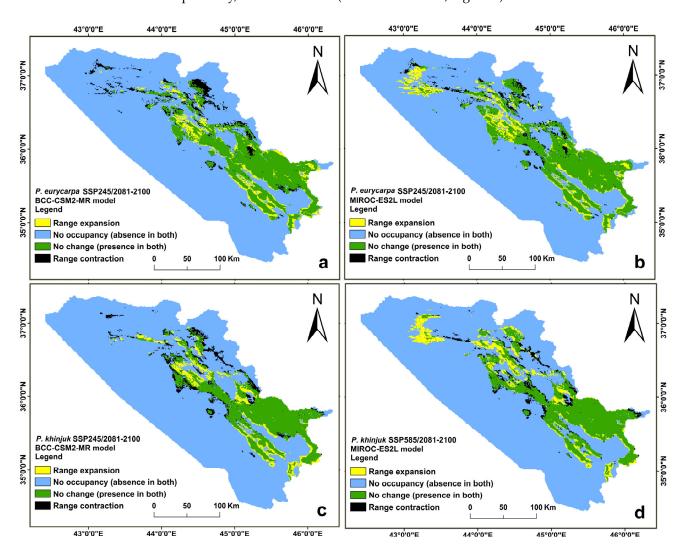
**Figure 6.** Distributions of the future habitat of *P. khinjuk* under (SSP 6585, 245, and 12) climate change scenarios in the BCC-CSM2-MR model for the years 2041–2060, 2061–2080, and 2081–2100 in the KRI.



**Figure 7.** Distributions of the future habitat of *P. khinjuk* under (SSP 585, 245, and 126) climate change scenarios in the MIROC-ES2L model for the years 2041–2060, 2061–2080, and 2081–2100 in the KRI.

# 3.3. Analysis of the Distribution Change between the Present and Future Habitats for P. khinjuk and P. eurycarpa

For both species, the habitat ranges were shown to undergo geographic distributional change (i.e., both expansion and contraction), as a result of climatic changes in the time frames 2041–2060, 2061–2080, and 2081–2100, using MIROC-ES2L and BCC-CSM2-MR models. Both models demonstrated that the total expansion magnitude was larger than the total contraction magnitude. The average expansion range for *P. eurycarpa* will be 1929 km² (3.776%) and 1793 km² (3.508%), under all the scenarios used. Additionally, for the two models, the average shrinkage is predicted to be 1304 km² (2.553%) and 1657 km² (3.244%), respectively (Tables A1 and A2). The average expansion range for *P. khinjuk* will be 2084 km² (4.079%) and 1783 km² (3.49%), for all of the applicable scenarios, whereas the average contraction is anticipated to be 1354 km² (2.65%) and 1698 km² (3.325%), respectively, for both models (Tables A3 and A4; Figure 8).

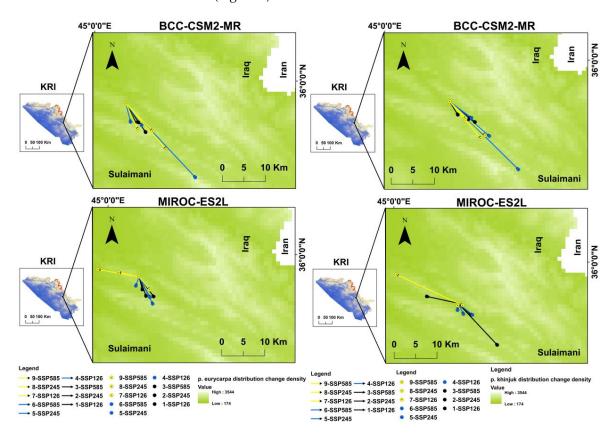


**Figure 8.** Change in the distribution of the habitat areas under the given scenarios of climate change for two species in future climates. Note: the maps illustrate only the expansion trend results for the species' habitat changes. (**a**,**b**) Figures show *P. eurycarpa*'s future expansion, while (**c**,**d**) figures show *P. khinjuk*'s future expansion.

### 3.4. The Direction and Degree of the Distributional Change for P. eurycarpa and P. khinjuk

The geographic coordinates for the two species are now at  $35^{\circ}91'40''$  north and  $45^{\circ}13'50''$  east, and  $35^{\circ}95'40''$  north and  $45^{\circ}06'40''$  east. Nevertheless, this centroid alters in the following ways under the scenarios of climate change: for *P. eurycarpa*, all of

the BCC-CSM2-MR scenarios predict a southeastward relocation of an average 8.51 km. This centroid, however, alters across the different climate change scenarios as follows: the majority of the MIROC-ES2L scenarios show a similar shift result toward the southeast, with the exception of SSP 585 (2061–2080), SSP 585 and SSP 245 (2081–2100), which are expected to shift  $45^{\circ}05'70''$  east and  $35^{\circ}93'60''$  north southwest,  $44^{\circ}97'80''$  east and  $35^{\circ}96'90''$  north northwest, and  $45^{\circ}02'20''$  east and  $35^{\circ}96'40''$  north northwest, respectively. However, *P. khinjuk*'s centroid change findings were not significantly different from *P. eurycarpa*'s. The centroid moved an average of 8.63 km toward the southeast, as predicted by the BCC-CSM2-MR scenarios. Except for SSP 126 (2041–2060), SSP 126 (2061–2080), and SSP 585 (2081–2100), which were relocated toward  $35^{\circ}92'90''$  north and  $45^{\circ}06'80''$  east northwest,  $35^{\circ}89'90''$  north and  $45^{\circ}13'30''$  east south, and  $35^{\circ}97'50''$  north and  $45^{\circ}00'40''$  east northwest, respectively, the majority of the MIROC-ES2L scenarios projected toward the southeast (Figure 9).



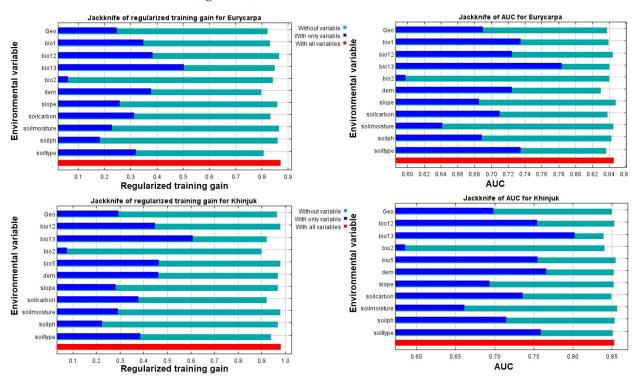
**Figure 9.** Change density (magnitude) of the distribution and the centroid (core) direction of *P. eurycarpa* (**left**) and *P. khinjuk* (**right**). The arrowhead = the distribution change direction under the given scenarios. The time windows of 2041–2060, 2061–2080, and 2081–2100 are shown in black, blue, and yellow color arrows in the figure, respectively.

3.5. Environmental Factors' Relative Relevance and Contribution to the Spread of P. eurycarpa and P. khinjuk

Four out of the five crucial factors that affect the likelihood of habitat distribution in the KRI are shared by both species. *P. eurycarpa* contributed to the following variables as follows: precipitation of the wettest month (Bio13) (32.2%), DEM (24.5%), soil type (16.2%), soil carbon (8.2%), and Geo (6.3%) (collectively 87.4). The least significant contributions were made by soil moisture, annual precipitation (Bio12), soil pH, temperature annual mean (Bio1), slope, and the range of mean diurnal (Bio2). Meanwhile, for *P. khinjuk*, the most significant variables were as follows: precipitation of the wettest month (Bio13) (35.2%), DEM (16.8%), soil type (i.e., Chromic Vertisols, lithosols, and Calcic Xerosols) (16.7%), the mean diurnal range (Bio2) (11.7%), and soil carbon (10.4%) (collectively 90.8). The

maximum temperature of the hottest month (Bio5), the annual precipitation (Bio12), soil pH, soil moisture, slope, and geo made the lowest relative contributions.

The MaxEnt model illustrated the relative impact of factors from the jackknife analysis on the two species under the present conditions. According to the jackknife test for regularizing the AUC variable gains and training gains (%), the precipitation of the soil carbon, soil type, annual precipitation (Bio12), DEM, annual mean temperature (Bio1), and wettest month (Bio13) included more information (gains) in the *P. eurycarpa* distribution compared to the rest of variables. For *P. khinjuk*, the variables soil carbon, soil type, and DEM included larger increases in distribution than the wettest month (Bio13), the annual precipitation(Bio12), the maximum temperature of the hottest month (Bio5), and the other factors (Figure 10).



**Figure 10.** The jackknife test for environmental factors significance, the AUC metric gain (%), and regularizing training gain (%) for *P. eurycarpa* (**top**) and *P. khinjuk* (**bottom**).

#### 4. Discussion

#### 4.1. Species of Tree Plant Respond Differently to the Scenarios of the Future Climate

One of the most urgent environmental challenges confronting mankind at the moment is global climate change. The average worldwide surface air temperature rose by 0.76 °C between 1850 and 2005. This indicates a trend of warming in the past 50 years of 0.13 °C per decade [28]. Between 2030 and 2052, if global warming keeps at its present pace, there will likely be a 1.5 °C increase [64]. Many plants are anticipated to adapt to the rapid temperature rise and spread to new areas. Climatic change has had a substantial impact on species distribution and abundance during the last several decades, and under future climate scenarios, it also causes species extinction, a shift in demography, a shift in phenology, expansion, contraction, and differences in growth and productivity [65]. According to Monzón, et al. [66], species ranges related to climate change are likely to grow or contract, and some species may even need to relocate to take advantage of new environments. According to earlier research, species contraction and plant habitat loss were anticipated to occur in the KRI [67], Africa [68], and California [69]. Different plant species also react differently to the predicted climatic circumstances. Nevertheless, other research, has indicated that, in several countries, including the Marion Island [70], Spain [71], the southern United States [72], and Indonesia [73], plant species' habitats are expanding due

Sustainability 2023, 15, 5469 14 of 25

to anticipated climate change. Our results also indicate that under the two GCMs, the BCC-CSM2-MR and the MIROC-ES2L models, and for the majority of the SSP climate change scenarios, the overall habit range of both species would expand. For the two GCMs the species habitat range would overlap by 8812 km<sup>2</sup> (17.25%) and 9177 km<sup>2</sup> (17.96%), respectively. The modeling revealed a total habitat suitability rise over a habitat suitability decline in the future circumstances. The modeling of BCC-CSM2-MR, for instance, under scenarios of SSP 245/2041-2060 and SSP 585/2041-2060, presented the maximum gain in habitat suitability for *P. khinjuk* and *P. eurycarpa*. Both species' habitat appropriateness is predicted to move from the present 9230 km<sup>2</sup> (18.06%) to the future 9735 km<sup>2</sup> (19.05%), and from the present  $10,442 \text{ km}^2$  (20.43%) to the future  $11,047 \text{ km}^2$  (21.61%), respectively. In a similar way, the scenarios of the MIROC-ES2L modeling for the two species resulted in a comparative finding regarding habitat suitability gain. P. khinjuk is predicted to expand from the current  $9230 \text{ km}^2$  (18.06%) to  $10,769 \text{ km}^2$  (21.07%) under SSP 585/2081-2100, in the future; whereas, P. eurycarpa, under SSP 585/2081-2100 was seen to move from the current 10,442 km<sup>2</sup> (20.43%) to 11,210 km<sup>2</sup> (21.94%) in the future (Tables A1 and A2). Moreover, the expansion of eminent habitat (the expansion-contraction net percentage) findings are predicted to increase in both modeling scenarios, MIROC-ES2L and BCC-CSM2-MR. The expansion in BCC-CSM2-MR will be as follows: P. khinjuk expands by 504 km<sup>2</sup> (0.986%) under SSP 245/2041-2060 and *P. eurycarpa* grows by 606 km<sup>2</sup> under SSP 585/2041-2060 (1.186%); whereas, P. khinjuk expands by 1539 km<sup>2</sup> (3.012%) under SSP 585/2081–2100 and *P. eurycarpa* expands by 1683 km<sup>2</sup> (3.294%) under SSP 245/2081–2100, when using MIROC-ES2L modeling scenarios (Tables A3 and A4).

Kozhoridze et al. [1] showed the same findings for the expansion of both species of *Pistacia*, in which they investigated *Pistacia* globally under future climate conditions between 2050 and 2100. Similarly, the eucalyptus spread in the 2050 and 2070 periods has been predicted by López-Sánchez, Castedo-Dorado, Cámara-Obregón and Barrio-Anta [71]. With regards to the tables resulting from the applied models and according to the centroid and distribution shift maps, for the *Pistacia* modeling in the KRI, utilizing MIROC-ES2L seems to be more appropriate compared to BCC-CSM2-MR. The MIROC-ES2L model predicts a greater growth and distribution range for the both species under the scenarios of the future climate. Moreover, the model indicated a considerable variety in the centroid shifts, which fits with the geography and the habitat characteristics of the research region. Moreover, according to the MIROC-ES2 scenarios, the future climate changes will insignificantly influence the amount of precipitation in the wettest months, which makes the greatest contribution to the distribution of both species in the current and future conditions (Table 2).

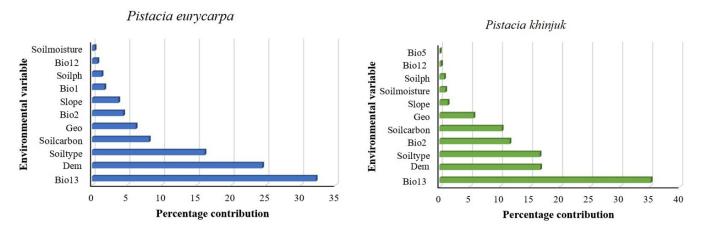
**Table 2.** The wettest month's precipitation varies depending on the present and the future conditions. Only the scenarios with the biggest growth ranges for the two species were selected.

		P. khinjuk		P. eurycarpa	
Precipitation of Wettest Month (Bio13)	Current	BCC-CSM2-MR SSP 245 (2041–2060)	MIROC-ES2L SSP 585 (2081–2100)	BCC-CSM2-MR SSP 585 (2041–2060)	MIROC-ES2L SSP 245 (2081–2100)
Lowest	58	55	56	55	55
Highest	200	174	198	181	193

4.2. Edaphic and Environmental Variables Contributed to the Expansion and Distribution of the P. khinjuk and P. eurycarpa in the Future Climatic Scenarios

The major environmental factors that affected both species are the precipitation in the wettest month (Bio13) (32.2%; 35.2%), DEM (24.5%; 16.8%), and soil type (16.2%; 16.7%) (Figure 11). These outcomes were anticipated and justified in light of the research area's geography, the seasonal change in precipitation, and the kind of soil. In the mountains of Zagros in the KRI, the two species have evolved to live on sandstone hills, steep dry slopes, and stony hillsides. The Zagros Mountains are characterized by significant seasonal

variations in temperature and precipitation, with dry summers and wet winters. The mountains of the KRI might get 1200 mm of precipitation annually, the most of which falls from October to May, the wettest months [40]. According to predictions on future climatic circumstances, the eco-regional features of the research area will make it ideal for the spread and growth of Pistacia. Additionally, P. khinjuk's strong resistance to osmotic stress and drought, which provide the plant with more moisture and regulate temperature and precipitation rates, improve the plants' ability to follow climatic change and survive aridity, according to [28]. As the most significant variable amongst other the variables, the results of Bio13 confirm the research conclusions. The predictions for Bio13 (i.e., the extracted layers of the map), which were employed in this work, are not anticipated to vary much in future climatic scenarios (especially in the mountainous regions of the KRI). The present range of the wettest month's anticipated precipitation is between 58 mm (the lowest) to 200 mm (the highest). Pistacia can effectively monitor climate change in the area since these values are not anticipated to vary considerably in the future climatic scenarios (Table 2). Despite these realities, P. khinjuk is expected to spread further in the mountainous region due to climate change. This indicates how the species' high tolerance aids the plant's ability to withstand summertime heat and aridity on the Zagros Mountains' stony slopes. Such findings support [28] findings that P. khinjuk grows better in high-altitude and dry environments. Similar justifications apply to *P. eurycarpa* as well, for two reasons: (i) both species share four out of five of the most significant variables that alter their potential distribution (Figure 11); and (ii) they usually coexist in most of the environments investigated in the field study (Figure 1). An additional factor that considerably influenced the spread of these two species was elevation. Elevation (or height) has been identified in several studies as a crucial element influencing the dispersion and growth of plants [74,75]. A piece of Marion Island research stated that 18 plant species reportedly expanded their ranges by about 70 m in altitude between 1966 and 2006, as a reaction to rising temperatures [70]. Moreover, the third factor influencing the dispersion of Pistacia was soil type. The distribution of the two species was influenced most by Chromic Vertisols, lithosols, and Calcic Xerosols, among the 19 categories of soil type. These soil types cover a region of around 23,021 km<sup>2</sup> (45%) in the overall research area, according to the retrieved soil type map layer (Table 3). These soil types are described in the FAO classification system as follows: Vertisols are swelling and shrinking clayey soils which occur in (sub) tropical areas with an expressed dry season and are dominated by expanding 2:1 lattice clays. The very shallow lithosols occur in the most eroding positions of the landscape and are less than 10 cm thick. While, the Xerosols are characterized by an aridic soil (having an aridic but not cold climate) and moderate contents of organic matter in the topsoil [76].



**Figure 11.** Contribution of the variables to the distribution probability of both *P. khinjuk* (**right**) and *P. eurycarpa* (**left**) in the KRI.

Table 3. Distribution of soil types by percentage in the KRI study area.

Value	Legend	FAOSOIL	Area	Area in %
1	Lithosol	I-Rc-Xk-c	5842.2	11.4
3	Lithosol	I-Be-c	515.4	1.0
4	Chromic Luvisols	Lc63-3bc	11.9	0.0
6	Calcic Xerosols	Xk5-3ab	101.1	0.2
7	Haplic Xerosols	Xh31-3a	30.7	0.1
9	Chromic Vertisols	Vc1-3a	9547	18.7
11	Haplic Xerosols	Xh33-3a	29.7	0.1
12	Chromic Vertisols	Vc50-3ab	2041	4.0
14	Lithosol	I-E-bc	8516	16.7
16	Lithosol	I-E-Xk-bc	5331	10.4
18	Calcic Xerosols	Xk29-ab	440	0.9
19	Calcic Xerosols	Xk26-2/3a	1065	2.1
20	Gypsic Xerosols	Xy5-a	1023	2.0
21	Calcic Xerosols	Xk28-b	12,464	24.4
22	Calcic Xerosols	Xk9-2/3a	586	1.1
23	Calcaric Fluvisols	Jc1-2a	462	0.9
24	Gypsic Yermosols	Yy10-2ab	259	0.5
25	Gypsic Yermosols	Yy10-2/3a	2473	4.8
29	Calcic Yermosols	Yk34-b	359	0.7

To analyze the environmental drivers of species distribution and to project the species' realized niche in a geographic area, SDMs can be utilized. The SDMs accuracy is affected by several predictors: (i) biological: presence or absence of the species; (ii) climatic: climate and vegetation; (iii) geological: topography, land cover types, soil and geology; and (iv) human activities: agriculture, deforestation, and land cover damage [77]. Therefore, researchers tend to consider habitat selection and other factors that influence species distribution [78]. Important limiters of SDMs involve the type of environmental gradients and multiple predictors, such as: the physical environment [79], a lack of biologically relevant information [80], invasive species and biotic interactions [81], and insufficient data [82]. In addition, among other factors that cause uncertainty in SDM output are data deficiencies (e.g., small and biased samples of species occurrences) and errors in the specification of the model [30]. Environmental gradients are among the other factors that influence SDM results, for example, direct gradients which have a direct physiological effect on species growth (temperature, pH, water, and nutrients), and indirect gradients, which result from location-specific correlations (elevation and latitude) [83].

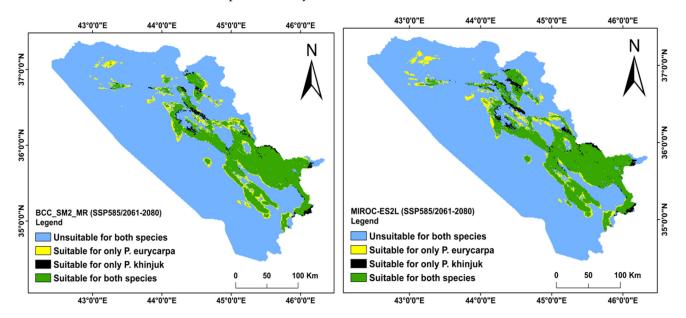
To minimize the uncertainties in SDM outputs, categorizing the habitat suitability of the studied area might be useful. Land suitability provides decision makers with a good insight to develop management strategy. In this study, habitat suitability was categorized into several layers, as follows: high suitability, medium suitability, low suitability, and unsuitable. These categorized layers represent how the climatic and edaphic factors influence the absence and presence of the species.

#### 4.3. Implications for Ecological Conservation

Through the protection and management of biological resources, conservation policy aims to provide a sustainable environment for the diverse range of species found across the globe without seriously compromising important habitats and ecosystems. Distribution, collection, storage, propagation, assessment, characterization, disease indexing, and disease removal are all part of conservation [84,85]. In the present and future climate situations, for conservation reasons, it is crucial to be aware of the possible range of a species. Therefore, some technologies (including global positioning systems (GPS), geographic information systems (GIS), SDMs, sampling schemes, and measurement methodologies are outstanding instruments to analyze the spatio-temporal dynamics and the complex processes that govern biodiversity. The vast array of remote sensing satellite data with their various

spatio-temporal resolution changes might be beneficial for biodiversity assessment [86]. Geospatial methods and ecological modeling are used in this study for forecasting and estimating the species distribution ranges. In particular, the incorporation of MaxEnt into the GIS environment has a significant function in enhancing conservation planning development and the input data [87].

Approximately 10,219 km² (20%) of the whole research area is predicted to be favorable for the distribution and development of the two species, according to the majority of the models' findings. The results of the modeling map indicate that the two species' medium and high suitability habitats for distribution are found in the mountainous area. The ecological, medicinal, nutritional, and cultural relevance of the *Pistacia* species, as well as their rarity in high altitudinal areas might benefit the KRI's forestry management. Additionally, the attention of environmentalists and conservationists should be encouraged to monitor climate change and help withstand prolonged dry spells. Consequently, the following are firmly proposed: (i) *Pistacia* has to be preserved and protected, especially in the overlapping areas 9177 km² (17.96%) for the MIROC-ES2L and 8812 km² (17.25%) for the BCC-CSM2-MR models (Figure 12), from anthropogenic activities, including burning, cutting and the overuse of plants; and (ii) these species depend on rocky mountain peaks and slope habitats (mostly the unchanged suitable areas), which are crucial ecosystems and should be prioritized by conservation efforts.



**Figure 12.** The two figures depict the spatial representation of the average overlapping habitat ranges among the 18 overlap results for *P. khinjuk* and *P. eurycarpa* in the KRI.

Regarding GPS data point collections and the plant specimen, this research had certain limitations. The biggest limitation was the problem of accessibility in some locations, especially in the east and north border areas where landmines have been planted or political tensions existed.

#### 5. Conclusions

The two *Pistacia* species' existing and future distributions in the KRI were estimated as part of this study's goal, which also included developing forest management strategies. Under the present environmental conditions, tree plants of *P. khinjuk* and *P. eurycarpa* cover around one fifth of the whole research area. The species attracted the attention of this study due to its cultural, medicinal, and ecological significance, and its tolerance to living in the rocky and mountainous highlands. The steep and rocky slopes of the Kurdistan Mountains are ideal sites for the development and spread of *Pistacia* species. The three main variables that affected these two species distributions were precipitation during the wettest

month (Bio13), DEM, and soil type. The entire habitat range of the species is anticipated to grow in the future scenarios of climate change, and the centroid is most likely going to migrate to the southeast. The highlands and overlapping ecosystems should be given priority in management plans and conservation measures. Thus, the classified present and prospective species distribution maps that have been developed in this work, especially the overlapping maps, provide invaluable planning insights for safeguarding the *Pistacia* species in the area. No prior research has modeled the geographic distribution of these two species under the existing and future climates in the KRI, according to the already available information. Geospatial techniques combined with correlation-based modeling are efficient tools for predicting the spatial patterns of the tree species in the mountain ecosystem. The expanded future habitat ranges particularly the overlapping areas, preferably between 1500 and 3500 masl., should be prioritized in conservation and management actions.

**Author Contributions:** N.R.K. conceived, designed the study; B.A.H. performed field surveys, data collection, analysis and writing; N.R.K. also participated in analysis, writing and compiling the final version of the manuscript. All authors have read and agreed to the published version of the manuscript.

Funding: This research received no external funding.

Institutional Review Board Statement: Not applicable.

Informed Consent Statement: Not applicable.

Data Availability Statement: Applicable on reasonable request.

**Acknowledgments:** The Kurdistan Botanical Foundation members helped the authors confirm the species identification of the specimens, for which they are thankful. We also want to express our gratitude to the local guides who shared important knowledge about the distribution, traditional uses, and morphology of the plants.

Conflicts of Interest: The authors declare no conflict of interest.

#### Appendix A





**Figure A1.** Herbarium samples of *P. khinjuk* (**left**) and *P. eurycarpa* (**right**), collected and prepared by the researcher Berhem A. HamadAmin.

**Table A1.** Changes in the distribution of the habitat range (areas) for *P. eurycarpa* in the present and future under the scenarios of climate change, BCC-CSM2-MR, SSP 585, 245, and 126 for 2041–60, 2061–80, and 2081–2100, respectively. The total area of the study equals 51,097 km<sup>2</sup> which is also shown in percentage below each category.

Year	Class	Range Expansion	No Occupancy (Absence in Both)	No Change (Presence in Both)	Range Contraction
	SSP 126 area (km <sup>2</sup> )	1659	38,996	9050	1392
	%	3.26	76.32	17.70	2.72
(2041–2060)	SSP 245 area (km <sup>2</sup> )	1724	38,931	9128	1314
(2041–2060)	%	3.38	76.20	17.86	2.56
	SSP 585 area (km <sup>2</sup> )	1893	38,762	9157	1285
	%	3.71	75.87	17.91	2.51
	SSP 126 area (km <sup>2</sup> )	1978	38,677	8559	1883
	%	3.88	75.70	16.74	3.68
(20(1, 2000)	SSP 245 area (km <sup>2</sup> )	1933	38,722	8057	2385
(2061–2080)	%	3.79	75.79	15.76	4.66
	SSP 585 area (km <sup>2</sup> )	1718	38,937	9080	1362
	%	3.37	76.21	17.76	2.66
	SSP 126 area (km <sup>2</sup> )	1637	39,018	8993	1449
	%	3.21	76.37	17.59	2.83
(2081–2100)	SSP 245 area (km²)	2111	38,544	8763	1679
(2001–2100)	%	4.14	75.44	17.14	3.28
	SSP 585 area (km <sup>2</sup> )	1489	39,166	8278	2164
	%	2.92	76.66	16.19	4.23

**Table A2.** Changes in the distribution of the habitat areas (range) for *P. eurycarpa* in the present and future under scenarios of climate change, MIROC-ES2L, SSP 585, 245, and 126 for 2041–60, 2061–80, and 2081–2100, respectively. The total area of the study equals 51,097 km<sup>2</sup> which is also shown in percentage below each category.

Year	Class	Range Expansion	No Occupancy (Absence in Both)	No Change (Presence in Both)	Range Contraction
	SSP 126 area (km <sup>2</sup> )	1658	38,997	8916	1526
	%	3.25	76.33	17.44	2.98
(2041–2060)	SSP 245 area (km <sup>2</sup> )	1744	38,911	9030	1412
(2041-2000)	%	3.42	76.16	17.66	2.76
	SSP 585 area (km <sup>2</sup> )	1556	39,099	9135	1307
	%	3.05	76.53	17.87	2.55
	SSP 126 area (km <sup>2</sup> )	1855	38,800	9169	1273
	%	3.64	75.94	17.94	2.48
(20(1 2000)	SSP 245 area (km <sup>2</sup> )	2045	38,610	9023	1421
(2061–2080)	%	4.01	75.57	17.65	2.77
	SSP 585 area (km <sup>2</sup> )	1897	38,758	9268	1174
	%	3.72	75.86	18.13	2.29
	SSP 126 area (km <sup>2</sup> )	1753	38,901	9238	1205
	%	3.44	76.14	18.07	2.35
(2001 2100)	SSP 245 area (km <sup>2</sup> )	2845	37,810	9283	1160
(2081–2100)	%	5.58	74	18.16	2.26
	SSP 585 area (km <sup>2</sup> )	2025	38,630	9185	1257
	%	3.97	75.61	17.97	2.45

**Table A3.** Changes in the distribution of the habitat areas (range) for *P. khinjuk* in the present and future under scenarios of climate change, BCC-CSM2-MR, SSP 585, 245, and 126 for 2041–60, 2061–80, and 2081–2100, respectively. The total area of the study equals 51,097 km<sup>2</sup> which is also shown in percentage below each category.

Year	Class	Range Expansion	No Occupancy (Absence in Both)	No Change (Presence in Both)	Range Contraction
	SSP 126 area (km <sup>2</sup> )	2044	39,824	7536	1693
	%	4.01	77.95	14.74	3.31
(2041–2060)	SSP 245 area (km <sup>2</sup> )	1833	40,035	7902	1327
(2041-2000)	%	3.60	78.36	15.46	2.59
	SSP 585 area (km²)	1913	39,955	7771	1458
	%	3.75	78.20	15.20	2.85
	SSP 126 area (km <sup>2</sup> )	1551	40,317	7437	1792
	%	3.04	78.91	14.55	3.50
(2061 2000)	SSP 245 area (km <sup>2</sup> )	1762	40,106	6727	2502
(2061–2080)	%	3.46	78.50	13.16	4.89
	SSP 585 area (km <sup>2</sup> )	1666	40,202	7750	1479
	%	3.27	78.69	15.16	2.89
	SSP 126 area (km <sup>2</sup> )	1786	40,082	7741	1488
	%	3.50	78.45	15.14	2.90
(2021 2100)	SSP 245 area (km <sup>2</sup> )	2075	39,793	7544	1685
(2081–2100)	%	4.07	77.89	14.76	3.29
	SSP 585 area (km <sup>2</sup> )	1431	40,437	7370	1859
	%	2.81	79.15	14.42	3.63

**Table A4.** Changes in the distribution of the habitat areas (range) for *P. khinjuk* in the present and future under scenarios of climate change, MIROC-ES2L, SSP 585, 245, and 126 for 2041–60, 2061–80, and 2081–2100, respectively. The total area of the study equals 51,097 km<sup>2</sup> which is also shown in percentage below each category.

Year	Class	Range Expansion	No Occupancy (Absence in Both)	No Change (Presence in Both)	Range Contraction
	SSP 126 area (km <sup>2</sup> )	2241	39,627	7876	1353
	%	4.39	77.56	15.41	2.64
(2041–2060)	SSP 245 area (km <sup>2</sup> )	1533	40,335	7511	1718
(2041-2000)	%	3.01	78.95	14.69	3.35
	SSP 585 area (km <sup>2</sup> )	1780	40,088	7909	1320
	%	3.49	78.46	15.47	2.58
	SSP 126 area (km <sup>2</sup> )	1891	39,977	7902	1327
	%	3.71	78.25	15.46	2.59
(20(1, 2000)	SSP 245 area (km <sup>2</sup> )	2154	39,714	7857	1372
(2061–2080)	%	4.22	77.73	15.37	2.68
	SSP 585 area (km <sup>2</sup> )	2002	39,867	7985	1243
	%	3.93	78.03	15.62	2.42
	SSP 126 area (km <sup>2</sup> )	1883	39,985	8054	1175
	%	3.69	78.26	15.75	2.29
(2001 2100)	SSP 245 area (km <sup>2</sup> )	2445	39,423	7853	1376
(2081–2100)	%	4.79	77.16	15.36	2.68
	SSP 585 area (km <sup>2</sup> )	2839	39,029	7931	1298
	%	5.56	76.39	15.51	2.53

Sustainability **2023**, 15, 5469 21 of 25

**Table A5.** Modeled habitat of the present and future areas of unsuitability and suitability distribution (in percentage) using model BCC-CSM2-MR under scenarios SSP 585, 245, and 126 climate change for *P. eurycarpa* for the years 2041–60, 2061–80, and 2081–2100, respectively. The total area of the study equals 51,097 km<sup>2</sup> which is also shown in percentage below each category.

Year	Class	Unsuitable	Low Suitability	Medium Suitability	High Suitability
<i>C</i> ,	Area (km²)	40,656	5032	4203	1208
Current	%	79.57	9.86	8.23	2.38
	SSP 126 area (km²)	40,391	5270	4469	969
	%	79.06	10.32	8.75	1.89
(2041–2060)	SSP 245 area (km <sup>2</sup> )	43,984	1538	4623	954
(2041-2060)	%	86.09	3.02	9.06	1.86
	SSP 585 area (km <sup>2</sup> )	40,051	5206	4902	940
	%	78.39	10.20	9.60	1.83
	SSP 126 area (km <sup>2</sup> )	40,566	4922	4496	1115
	%	79.40	9.64	8.81	2.17
(20(1, 2000)	SSP 245 area (km <sup>2</sup> )	41,109	4141	4820	1029
(2061–2080)	%	80.46	8.11	9.42	2.01
	SSP 585 area (km <sup>2</sup> )	40,299	5349	4555	896
	%	78.88	10.48	8.92	1.75
	SSP 126 area (km²)	40,468	5146	4589	896
	%	79.21	10.08	8.99	1.75
(2001 2100)	SSP 245 area (km <sup>2</sup> )	40,226	4676	5163	1034
(2081–2100)	%	78.73	9.16	10.11	2.02
	SSP 585 area (km <sup>2</sup> )	41,330	5000	3868	901
	%	80.89	9.79	7.58	1.76

**Table A6.** Modeled habitat of the present and future areas of unsuitability and suitability distribution (in percentage) under SSP 585, 245, and 126 climate change scenarios using the MIROC-ES2L model for *P. eurycarpa* for the years 2041–60, 2061–80, and 2081–2100, respectively. The total area of the study equals 51,097 km<sup>2</sup> which is also shown in percentage below each category.

Year	Class	Unsuitable	Low Suitability	Medium Suitability	High Suitability
Comment.	Area (km²)	40,656	5032	4201	1208
Current	%	79.57	9.86	8.21	2.36
	SSP 126 area (km <sup>2</sup> )	40,523	5220	4415	939
	%	79.31	10.22	8.63	1.83
(2041–2060)	SSP 245 area (km²)	40,325	5347	4507	918
(2041–2000)	%	78.93	10.47	8.81	1.79
	SSP 585 area (km²)	40,408	5357	4427	905
	%	79.09	10.49	8.66	1.76
	SSP 126 area (km <sup>2</sup> )	40,073	5469	4703	852
	%	78.43	10.71	9.20	1.66
(2061–2080)	SSP 245 area (km²)	40,031	5217	4859	990
(2001–2000)	%	78.35	10.22	9.50	1.93
	SSP 585 area (km <sup>2</sup> )	39,932	5454	4777	934
	%	78.16	10.68	9.34	1.82
	SSP 126 area (km <sup>2</sup> )	40,107	5429	4628	933
	%	78.50	10.63	9.05	1.82
(2001 2100)	SSP 245 area (km <sup>2</sup> )	38,974	5521	5397	1205
(2081–2100)	%	76.28	10.81	10.55	2.35
	SSP 585 area (km <sup>2</sup> )	39,888	5693	4547	969
	%	78.07	11.15	8.89	1.89

**Table A7.** Modeled habitat of the present and future areas of unsuitability and suitability distribution (in percentage) under SSP 585, 245, and 126 climate change scenarios using the BCC-CSM2-MR model for *P. khinjuk* for the years 2041–60, 2061–80, and 2081–2100, respectively. The total area of the study equals 51,097 km<sup>2</sup> which is also shown in percentage below each category.

Year	Class	Unsuitable	Low Suitability	Medium Suitability	High Suitability
C	Area (km²)	41,868	4308	3771	1150
Current	%	81.95	8.44	7.37	2.24
	SSP 126 area (km <sup>2</sup> )	41,518	4811	4153	615
	%	81.26	9.42	8.12	1.20
(2041, 2060)	SSP 245 area (km²)	41,363	5048	4084	602
(2041–2060)	%	80.96	9.89	8	1.19
	SSP 585 area (km <sup>2</sup> )	41,420	4808	4201	668
	%	81.07	9.42	8.21	1.30
	SSP 126 area (km <sup>2</sup> )	42,111	4269	3905	812
	%	82.42	8.36	7.63	1.58
(20(1, 2000)	SSP 245 area (km²)	42,613	3854	3819	811
(2061–2080)	%	83.40	7.55	7.47	1.58
	SSP 585 area (km <sup>2</sup> )	41,682	4838	3942	635
	%	81.58	9.48	7.71	1.23
	SSP 126 area (km <sup>2</sup> )	41,575	4700	4211	611
	%	81.37	9.21	8.23	1.21
(2081–2100)	SSP 245 area (km <sup>2</sup> )	41,484	4610	4315	688
(2001–2100)	%	81.19	9.03	8.44	1.34
	SSP 585 area (km <sup>2</sup> )	42,298	4331	3714	754
	%	82.79	8.48	7.26	1.47

**Table A8.** Modeled habitat of the present and future areas of unsuitability and suitability distribution (in percentage under SSP 585, 245, and 126 climate change scenarios) using the MIROC-ES2L model for *P. khinjuk* for the following years 2041–60, 2061–80, and 2081–2100, respectively. The total area of the study equals 51,097 km<sup>2</sup> which is also shown in percentage below each category.

Year	Class	Unsuitable	Low Suitability	Medium Suitability	High Suitability
	Area (km²)	41,868	4308	3771	1150
Current	%	81.95	8.44	7.37	2.24
	SSP 126 area (km <sup>2</sup> )	40,984	4568	4204	1341
	%	80.22	8.95	8.22	2.62
(2041–2060)	SSP 245 area (km <sup>2</sup> )	42,060	4159	4108	770
(2041-2060)	%	82.32	8.15	8.03	1.50
	SSP 585 area (km <sup>2</sup> )	41,409	4944	4038	706
	%	81.05	9.68	7.89	1.37
	SSP 126 area (km <sup>2</sup> )	41,305	4920	4306	566
	%	80.84	9.64	8.42	1.10
(2061 2080)	SSP 245 area (km <sup>2</sup> )	41,087	4975	4445	590
(2061–2080)	%	80.42	9.74	8.69	1.15
	SSP 585 area (km <sup>2</sup> )	41,110	5011	4181	795
	%	80.44	9.81	8.17	1.55
	SSP 126 area (km <sup>2</sup> )	41,163	4987	4224	723
	%	80.57	9.77	8.26	1.41
(2001 2100)	SSP 245 area (km <sup>2</sup> )	40,800	5524	4169	604
(2081–2100)	%	79.86	10.82	8.15	1.17
	SSP 585 area (km <sup>2</sup> )	40,329	5407	4246	1115
	%	78.93	10.59	8.30	2.17

#### References

 Kozhoridze, G.; Orlovsky, N.; Orlovsky, L.; Blumberg, D.G.; Golan-Goldhirsh, A. Geographic distribution and migration pathways of *Pistacia*-present, past and future. *Ecography* 2015, 38, 1141–1154. [CrossRef]

- 2. AL-Saghir, M.G.; Porter, D.M. Taxonomic revision of the genus *Pistacia* L. (Anacardiaceae). *Am. J. Plant Sci.* **2011**, *3*, 12–32. [CrossRef]
- 3. Guest, E.; Townsend, C. Flora of Iraq; Ministry of Agriculture of the Republic of Iraq: Baghdad, Iraq, 1966.
- 4. Ahmed, H.M. Traditional uses of Kurdish medicinal plant *Pistacia atlantica* subsp. *kurdica* Zohary in Ranya, Southern Kurdistan. *Genet. Resour. Crop Evol.* **2017**, 64, 1473–1484. [CrossRef]
- 5. Rankou, H.; M'sou, S.; Babahmad, R.A.; Ouhammou, A.; Alifriqui, M.; Martin, G. Pistacia atlantica. In *The IUCN Red List of Threatened Species*; IUCN red list: Cambridge, UK, 2018.
- 6. Khwarahm, N.R. Spatial modeling of land use and land cover change in Sulaimani, Iraq, using multitemporal satellite data. *Environ. Monit. Assess.* **2021**, 193, 148. [CrossRef]
- 7. Khwarahm, N.R.; Najmaddin, P.M.; Ararat, K.; Qader, S. Past and future prediction of land cover land use change based on earth observation data by the CA–Markov model: A case study from Duhok governorate, Iraq. *Arab. J. Geosci.* **2021**, *14*, 1544. [CrossRef]
- 8. Khwarahm, N.R.; Qader, S.; Ararat, K.; Al-Quraishi, A.M.F. Predicting and mapping land cover/land use changes in Erbil/Iraq using CA-Markov synergy model. *Earth Sci. Inform.* **2021**, *14*, 393–406. [CrossRef]
- 9. Khwarahm, N.R. Modeling forest-shrubland fire susceptibility based on machine learning and geospatial approaches in mountains of Kurdistan Region, Iraq. *Arab. J. Geosci.* **2022**, *15*, 1184. [CrossRef]
- 10. Nasser, M. Forests and forestry in Iraq: Prospects and limitations. Commonw. For. Rev. 1984, 63, 299–304.
- 11. Javanshah, A. Global warming has been affecting some morphological characters of pistachio trees (*Pistacia vera* L.). *Afr. J. Agric. Res.* **2010**, *5*, 3394–3401.
- 12. Hama, A.A.; Khwarahm, N.R. Predictive mapping of two endemic oak tree species under climate change scenarios in a semiarid region: Range overlap and implications for conservation. *Ecol. Inform.* **2023**, *73*, 101930. [CrossRef]
- 13. Radha, K.O.; Khwarahm, N.R. An Integrated Approach to Map the Impact of Climate Change on the Distributions of *Crataegus azarolus* and *Crataegus monogyna* in Kurdistan Region, Iraq. *Sustainability* **2022**, *14*, 14621. [CrossRef]
- 14. Ahmed, Z.B.; Yousfi, M.; Viaene, J.; Dejaegher, B.; Demeyer, K.; Heyden, Y.V. Four *Pistacia atlantica* subspecies (*atlantica*, *cabulica*, *kurdica* and *mutica*): A review of their botany, ethnobotany, phytochemistry and pharmacology. *J. Ethnopharmacol.* **2020**, 265, 113329. [CrossRef] [PubMed]
- 15. Bozorgi, M.; Memariani, Z.; Mobli, M.; Salehi Surmaghi, M.H.; Shams-Ardekani, M.R.; Rahimi, R. Five *Pistacia* species (*P. vera, P. atlantica, P. terebinthus, P. khinjuk*, and *P. lentiscus*): A review of their traditional uses, phytochemistry, and pharmacology. *Sci. World J.* 2013, 2013, 219815. [CrossRef] [PubMed]
- 16. Hatamnia, A.A.; Abbaspour, N.; Darvishzadeh, R. Antioxidant activity and phenolic profile of different parts of Bene (*Pistacia atlantica* subsp. *kurdica*) fruits. *Food Chem.* **2014**, *145*, 306–311. [CrossRef] [PubMed]
- 17. Bahmani, M.; Saki, K.; Asadbeygi, M.; Adineh, A.; Saberianpour, S.; Rafieian-Kopaei, M.; Bahmani, F.; Bahmani, E. The effects of nutritional and medicinal mastic herb (*Pistacia atlantica*). *J. Chem. Pharm. Res.* **2015**, *7*, 646–653.
- 18. Schulze-Kaysers, N.; Feuereisen, M.; Schieber, A. Phenolic compounds in edible species of the Anacardiaceae family—A review. *RSC Adv.* **2015**, *5*, 73301–73314. [CrossRef]
- 19. Sharifi, M.S.; Hazell, S.L. GC-MS analysis and antimicrobial activity of the essential oil of the trunk exudates from *Pistacia atlantica kurdica*. *J. Pharm. Sci. Res.* **2011**, *3*, 1364.
- 20. Ahmad, S.A.; Askari, A.A. Ethnobotany of the Hawraman region of Kurdistan Iraq. Harv. Pap. Bot. 2015, 20, 85–89. [CrossRef]
- 21. Al-Saghir, M.G.; Porter, D.M.; Nilsen, E.T. Leaf anatomy of Pistacia species (Anacardiaceae). J. Biol. Sci. 2006, 6, 242–244. [CrossRef]
- 22. Basr Ila, H.; Kafkas, S.; Topaktas, M. Chromosome numbers of four *Pistacia* (Anacardiaceae) species. *J. Hortic. Sci. Biotechnol.* **2003**, 78, 35–38. [CrossRef]
- 23. Parfitt, D.E.; Badenes, M.L. Phylogeny of the genus *Pistacia* as determined from analysis of the chloroplast genome. *Proc. Natl. Acad. Sci. USA* **1997**, *94*, 7987–7992. [CrossRef] [PubMed]
- 24. Kafkas, S.; Perl-Treves, R. Morphological and molecular phylogeny of *Pistacia* species in Turkey. *Theor. Appl. Genet.* **2001**, 102, 908–915. [CrossRef]
- 25. Talebi, M.; Kazemi, M.; Sayed-Tabatabaei, B.E. Molecular diversity and phylogenetic relationships of *Pistacia vera*, *Pistacia atlantica* subsp. *mutica* and *Pistacia khinjuk* using SRAP markers. *Biochem. Syst. Ecol.* **2012**, 44, 179–185. [CrossRef]
- 26. Belhadj, S.; Derridj, A.; Aigouy, T.; Gers, C.; Gauquelin, T.; Mevy, J.P. Comparative morphology of leaf epidermis in eight populations of Atlas pistachio (*Pistacia atlantica* Desf., Anacardiaceae). *Microsc. Res. Tech.* **2007**, *70*, 837–846. [CrossRef]
- 27. Sawidis, T.; Dafnis, S.; Weryzko-Chmielewska, E. Distribution, development and structure of resin ducts in *Pistacia lentiscus* var. *chia* Duhamel. *Flora* **2000**, *195*, 83–94. [CrossRef]
- 28. Al-Alfy, N.; Moustafa, A.; Alotaibi, M.; Mansour, S. Impact of Climate Change on *Pistacia khinjuk* as a Medicinal Plant in Egypt and Saudi Arabia. *Appl. Sci. Res. Rev.* **2019**, *6*, 3.
- 29. Pineda, E.; Lobo, J.M. Assessing the accuracy of species distribution models to predict amphibian species richness patterns. *J. Anim. Ecol.* **2009**, *78*, 182–190. [CrossRef]
- 30. Elith, J.; Leathwick, J.R. Species distribution models: Ecological explanation and prediction across space and time. *Annu. Rev. Ecol. Evol. Syst.* **2009**, *40*, 677–697. [CrossRef]

Sustainability 2023, 15, 5469 24 of 25

31. Svenning, J.-C.; Fløjgaard, C.; Marske, K.A.; Nógues-Bravo, D.; Normand, S. Applications of species distribution modeling to paleobiology. *Quat. Sci. Rev.* **2011**, *30*, 2930–2947. [CrossRef]

- 32. Elith, J.; Graham, C.H.; Anderson, R.P.; Dudík, M.; Ferrier, S.; Guisan, A.; Hijmans, R.J.; Huettmann, F.; Leathwick, J.R.; Lehmann, A.; et al. Novel methods improve prediction of species' distributions from occurrence data. *Ecography* **2006**, 29, 129–151. [CrossRef]
- 33. Graham, C.H.; Ferrier, S.; Huettman, F.; Moritz, C.; Peterson, A.T. New developments in museum-based informatics and applications in biodiversity analysis. *Trends Ecol. Evol.* **2004**, *19*, 497–503. [CrossRef] [PubMed]
- 34. Phillips, S.J.; Anderson, R.P.; Schapire, R.E. Maximum entropy modeling of species geographic distributions. *Ecol. Model.* **2006**, 190, 231–259. [CrossRef]
- 35. Bor, N.L.; Guest, E. Flora of Iraq, Vol. 9. Gramineae. In Flora of Iraq; Ministry of Agriculture and Agrarian Reform: Baghdad, Iraq, 1968.
- 36. Malinowski, J.C. *Iraq: A Geography*; United States Military Academy: West Point, NY, USA, 2002.
- 37. Salman, S.A.; Shahid, S.; Ismail, T.; Ahmed, K.; Chung, E.-S.; Wang, X.-J. Characteristics of annual and seasonal trends of rainfall and temperature in Iraq. *Asia-Pac. J. Atmos. Sci.* **2019**, *55*, 429–438. [CrossRef]
- 38. Sissakian, V.; Jabbar, M.A.; Al-Ansari, N.; Knutsson, S. Development of Gulley Ali Beg Gorge in Rawandooz Area, Northern Iraq. *Engineering* **2015**, *7*, 16–30. [CrossRef]
- 39. Khwarahm, N.R.; Ararat, K.; Qader, S.; Sabir, D.K. Modeling the distribution of the Near Eastern fire salamander (*Salamandra infraimmaculata*) and Kurdistan newt (*Neurergus derjugini*) under current and future climate conditions in Iraq. *Ecol. Inform.* **2021**, 63, 101309. [CrossRef]
- 40. Gaznayee, H.A.A.; Al-Quraishi, A.M.F.; Mahdi, K.; Ritsema, C. A Geospatial Approach for Analysis of Drought Impacts on Vegetation Cover and Land Surface Temperature in the Kurdistan Region of Iraq. *Water* **2022**, *14*, 927. [CrossRef]
- 41. Bhatta, K.P.; Chaudhary, R.P.; Vetaas, O.R. A comparison of systematic versus stratified-random sampling design for gradient analyses: A case study in subalpine Himalaya, Nepal. *Phytocoenologia* **2012**, *42*, 191–202. [CrossRef]
- 42. Boakes, E.H.; McGowan, P.J.; Fuller, R.A.; Chang-qing, D.; Clark, N.E.; O'Connor, K.; Mace, G.M. Distorted views of biodiversity: Spatial and temporal bias in species occurrence data. *PLoS Biol.* **2010**, *8*, e1000385. [CrossRef]
- 43. Radosavljevic, A.; Anderson, R.P. Making better Maxent models of species distributions: Complexity, overfitting and evaluation. J. Biogeogr. 2014, 41, 629–643. [CrossRef]
- 44. Brown, J.L. SDM toolbox: A python-based GIS toolkit for landscape genetic, biogeographic and species distribution model analyses. *Methods Ecol. Evol.* **2014**, *5*, 694–700. [CrossRef]
- 45. Soberón, J.; Peterson, A.T. Interpretation of models of fundamental ecological niches and species' distributional areas. *Biodivers*. *Inform.* **2005**, 2, 1–10. [CrossRef]
- 46. Elith, J.; Kearney, M.; Phillips, S. The art of modelling range-shifting species. Methods Ecol. Evol. 2010, 1, 330–342. [CrossRef]
- 47. Austin, M.P.; Van Niel, K.P. Improving species distribution models for climate change studies: Variable selection and scale. *J. Biogeogr.* **2011**, *38*, 1–8. [CrossRef]
- 48. Pradervand, J.-N.; Dubuis, A.; Pellissier, L.; Guisan, A.; Randin, C. Very high resolution environmental predictors in species distribution models: Moving beyond topography? *Prog. Phys. Geogr.* **2014**, *38*, 79–96. [CrossRef]
- 49. IPCC. The Physical Science Basis: Contribution of Working Group I to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change; Cambridge University Press: Cambridge, UK; New York, NY, USA, 2007; Volume 996, pp. 113–119.
- 50. Hijmans, R.J.; Cameron, S.E.; Parra, J.L.; Jones, P.G.; Jarvis, A. Very high resolution interpolated climate surfaces for global land areas. *Int. J. Climatol. J. R. Meteorol. Soc.* **2005**, 25, 1965–1978. [CrossRef]
- 51. Hanberry, B.B. Global population densities, climate change, and the maximum monthly temperature threshold as a potential tipping point for high urban densities. *Ecol. Indic.* **2022**, *135*, 108512. [CrossRef]
- 52. Karim, R.; Tan, G.; Ayugi, B.; Babaousmail, H.; Liu, F. Evaluation of historical CMIP6 model simulations of seasonal mean temperature over Pakistan during 1970–2014. *Atmosphere* **2020**, *11*, 1005. [CrossRef]
- 53. Fan, X.; Duan, Q.; Shen, C.; Wu, Y.; Xing, C. Global surface air temperatures in CMIP6: Historical performance and future changes. *Environ. Res. Lett.* **2020**, *15*, 104056. [CrossRef]
- 54. Beck, H.E.; Zimmermann, N.E.; McVicar, T.R.; Vergopolan, N.; Berg, A.; Wood, E.F. Present and future Köppen-Geiger climate classification maps at 1-km resolution. *Sci. Data* **2018**, *5*, 180214. [CrossRef]
- 55. Syfert, M.M.; Smith, M.J.; Coomes, D.A. The effects of sampling bias and model complexity on the predictive performance of MaxEnt species distribution models. *PLoS ONE* **2013**, *8*, e55158. [CrossRef]
- 56. Brown, J.L.; Bennett, J.R.; French, C.M. SDMtoolbox 2.0: The next generation Python-based GIS toolkit for landscape genetic, biogeographic and species distribution model analyses. *PeerJ* 2017, 5, e4095. [CrossRef] [PubMed]
- 57. Dudík, M.; Phillips, S.J.; Schapire, R.E. Maximum entropy density estimation with generalized regularization and an application to species distribution modeling. *J. Mach. Learn. Res.* **2007**, *8*, 1217–1260.
- 58. Merow, C.; Smith, M.J.; Silander Jr, J.A. A practical guide to MaxEnt for modeling species' distributions: What it does, and why inputs and settings matter. *Ecography* **2013**, *36*, 1058–1069. [CrossRef]
- 59. Phillips, S.J. A brief tutorial on Maxent. ATT Res. 2005, 190, 231–259.
- 60. Jiménez-Valverde, A.; Lobo, J.M. Threshold criteria for conversion of probability of species presence to either–or presence–absence. *Acta Oecologica* **2007**, *31*, 361–369. [CrossRef]
- 61. Yang, X.-Q.; Kushwaha, S.; Saran, S.; Xu, J.; Roy, P. Maxent modeling for predicting the potential distribution of medicinal plant, Justicia adhatoda L. in Lesser Himalayan foothills. *Ecol. Eng.* **2013**, *51*, 83–87. [CrossRef]

Sustainability **2023**, 15, 5469 25 of 25

62. Jiang, H.; Liu, T.; Li, L.; Zhao, Y.; Pei, L.; Zhao, J. Predicting the potential distribution of *Polygala tenuifolia* Willd. under climate change in China. *PLoS ONE* **2016**, *11*, e0163718. [CrossRef]

- 63. Allouche, O.; Tsoar, A.; Kadmon, R. Assessing the accuracy of species distribution models: Prevalence, kappa and the true skill statistic (TSS). *J. Appl. Ecol.* **2006**, *43*, 1223–1232. [CrossRef]
- 64. Li, C.; Gao, Y.; Zhao, Z.; Ma, D.; Zhou, R.; Wang, J.; Zhang, Q.; Liu, Q. Potential geographical distribution of *Anopheles gambiae* worldwide under climate change. *J. Biosaf. Biosecur.* **2021**, *3*, 125–130. [CrossRef]
- 65. Hoveka, L.N.; van der Bank, M.; Davies, T.J. Winners and losers in a changing climate: How will protected areas conserve red list species under climate change? *Divers. Distrib.* **2022**, *28*, 782–792. [CrossRef]
- 66. Monzón, J.; Moyer-Horner, L.; Palamar, M.B. Climate change and species range dynamics in protected areas. *Bioscience* **2011**, 61, 752–761. [CrossRef]
- 67. Khwarahm, N.R. Mapping current and potential future distributions of the oak tree (*Quercus aegilops*) in the Kurdistan Region, Iraq. *Ecol. Process.* **2020**, *9*, 56. [CrossRef]
- 68. Ayebare, S.; Plumptre, A.; Kujirakwinja, D.; Segan, D. Conservation of the endemic species of the Albertine Rift under future climate change. *Biol. Conserv.* **2018**, 220, 67–75. [CrossRef]
- 69. Loarie, S.R.; Carter, B.E.; Hayhoe, K.; McMahon, S.; Moe, R.; Knight, C.A.; Ackerly, D.D. Climate change and the future of California's endemic flora. *PLoS ONE* **2008**, *3*, e2502. [CrossRef] [PubMed]
- 70. Le Roux, P.C.; McGEOCH, M.A. Rapid range expansion and community reorganization in response to warming. *Glob. Chang. Biol.* **2008**, *14*, 2950–2962. [CrossRef]
- 71. López-Sánchez, C.A.; Castedo-Dorado, F.; Cámara-Obregón, A.; Barrio-Anta, M. Distribution of *Eucalyptus globulus* Labill. in northern Spain: Contemporary cover, suitable habitat and potential expansion under climate change. *For. Ecol. Manag.* **2021**, 481, 118723. [CrossRef]
- 72. Wang, W.J.; Thompson III, F.R.; He, H.S.; Fraser, J.S.; Dijak, W.D.; Jones-Farrand, T. Climate change and tree harvest interact to affect future tree species distribution changes. *J. Ecol.* **2019**, *107*, 1901–1917. [CrossRef]
- 73. Setyawan, A.D.; Supriatna, J.; Nisyawati, N.; Nursamsi, I.; Sutarno, S.; Sugiyarto, S.; Sunarto, S.; Pradan, P.; Budiharta, S.; Pitoyo, A.; et al. Projecting expansion range of *Selaginella zollingeriana* in the Indonesian archipelago under future climate condition. *Biodivers. J. Biol. Divers.* 2021, 22, 2088–2103. [CrossRef]
- 74. Tsiftsis, S.; Štípková, Z.; Kindlmann, P. Role of way of life, latitude, elevation and climate on the richness and distribution of orchid species. *Biodivers. Conserv.* **2019**, *28*, 75–96. [CrossRef]
- 75. Zhang, K.; Liu, H.; Pan, H.; Shi, W.; Zhao, Y.; Li, S.; Liu, J.; Tao, J. Shifts in potential geographical distribution of *Pterocarya stenoptera* under climate change scenarios in China. *Ecol. Evol.* **2020**, *10*, 4828–4837. [CrossRef]
- 76. Nachtergaele, F.O. Classification Systems: FAO☆. In *Reference Module in Earth Systems and Environmental Sciences*; Elsevier: Amsterdam, The Netherlands, 2017.
- 77. Hao, T.; Elith, J.; Guillera-Arroita, G.; Lahoz-Monfort, J.J. A review of evidence about use and performance of species distribution modelling ensembles like BIOMOD. *Divers. Distrib.* **2019**, 25, 839–852. [CrossRef]
- 78. Baldwin, R.A. Use of maximum entropy modeling in wildlife research. Entropy 2009, 11, 854–866. [CrossRef]
- 79. Elith, J.; Franklin, J. Species distribution modeling. In *Encyclopedia of Biodiversity*, 2nd ed.; Elsevier Inc.: Amsterdam, The Netherlands, 2013; pp. 692–705.
- 80. Barnhart, P.R.; Gillam, E.H. The impact of sampling method on maximum entropy species distribution modeling for bats. *Acta Chiropterologica* **2014**, *16*, 241–248. [CrossRef]
- 81. Kearney, M.; Porter, W. Mechanistic niche modelling: Combining physiological and spatial data to predict species' ranges. *Ecol. Lett.* **2009**, 12, 334–350. [CrossRef] [PubMed]
- 82. Carneiro, L.R.d.A.; Lima, A.P.; Machado, R.B.; Magnusson, W.E. Limitations to the use of species-distribution models for environmental-impact assessments in the Amazon. *PLoS ONE* **2016**, *11*, e0146543. [CrossRef]
- 83. Miller, J. Species distribution modeling. Geogr. Compass 2010, 4, 490–509. [CrossRef]
- 84. Kasagana, V.N.; Karumuri, S.S. Conservation of medicinal plants (past, present & future trends). J. Pharm. Sci. Res. 2011, 3, 1378.
- 85. Mahmoodi, S.; Heydari, M.; Ahmadi, K.; Khwarahm, N.R.; Karami, O.; Almasieh, K.; Naderi, B.; Bernard, P.; Mosavi, A. The current and future potential geographical distribution of *Nepeta crispa* Willd., an endemic, rare and threatened aromatic plant of Iran: Implications for ecological conservation and restoration. *Ecol. Indic.* **2022**, *137*, 108752. [CrossRef]
- 86. Yadav, P.; Sarma, K.; Dookia, S. The review of biodiversity and conservation study in India using geospatial technology. *Int. J. Remote Sens. GIS* **2013**, *2*, 1–10.
- 87. Draper, D.; Rosselló-Graell, A.; Garcia, C.; Gomes, C.T.; Sérgio, C.l. Application of GIS in plant conservation programmes in Portugal. *Biol. Conserv.* **2003**, *113*, 337–349. [CrossRef]

**Disclaimer/Publisher's Note:** The statements, opinions and data contained in all publications are solely those of the individual author(s) and contributor(s) and not of MDPI and/or the editor(s). MDPI and/or the editor(s) disclaim responsibility for any injury to people or property resulting from any ideas, methods, instructions or products referred to in the content.